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**ADSORPTION OF Fe^{2+} AND Zn^{2+} IONS FROM LANDFILL
LEACHATE BY NATURAL BENTONITE FROM KRIVA PALANKA(B
- KP), REPUBLIC OF MACEDONIA**

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ABSTRACT

Due to its typical layered silicate structure, large specific surface area, chemical and mechanical stability, high cation exchange capacity, bentonite has been widely applied as heavy metals adsorbent from aqueous solutions. In this paper the adsorptive properties of natural bentonite from Kriva Palanka, Republic of Macedonia, for the removal of Fe^{2+} and Zn^{2+} ions from landfill leachate were studied. The raw material (B-KP) was characterized in terms of chemical composition (conventional silicate analysis), structural morphology (XRD, FT-IR, TGA-DTA), specific surface area, particle size distribution (wet - sieve analysis). It was considered that montmorillonite is dominant phase in B-KP. For adsorption experiments the batch method has been employed using volume of 25 cm³, 50 cm³, 75 cm³, 100 cm³ and 125 cm³ of leachate with 2 g of bentonite, at 298 K. The initial metal ion concentration in the leachate was 9,25 mg/dm³ for Fe^{2+} and 0,22 mg/dm³ for Zn^{2+} and these concentrations as well as the concentrations after adsorption were determined by atomic absorption spectrophotometer (AAS). Applied Freundlich and Langmuir adsorption isotherms showed good fits to the experimental data. The monolayer capacities of bentonite for Fe^{2+} and Zn^{2+} were determined by Langmuir equation and are 0,4371 mg/g and $4,1 \cdot 10^{-3}$ mg/g, respectively. The removal efficiency for zinc was above 98%.

Key words: municipal solid waste leachate, Fe^{2+} , Zn^{2+} , adsorption isotherms, bentonite.

INTRODUCTION

Due to the irregular management, municipal solid waste landfills (MSWL's) are major threat to the quality of the environment in Macedonia as in many other developing countries. Produced and freely released leachate is a source of groundwater and soil pollution. Leachate is highly polluted liquid composed of many inorganic (sulfates, chlorides, heavy metals) and organic (acids, alcohols) pollutants. The MSWL Drisla, does not contain any drainage system for the underground waters protection nor any receiving reservoir for primary treatment. The leachate is freely released into the stream of Meckin Dol. Therefore the presence of heavy metals in leachate is a serious threat for groundwater and nearby fields and accordingly for the health of the population. .

The methods used for removal of heavy metals include chemical precipitation, ion exchange, solvent extraction, membrane processes, adsorption and others. Among all the approaches proposed, adsorption is one of the most popular method and it is considered as an effective, efficient and economic method for wastewater purification. Various materials such as activated carbon (Kouakou et al., 2013; Bernard and Jimoh 2013) natural and synthetic zeolites (Erdem et al., 2004; Shaheen et al., 2012), bentonite (Ghormi et al., 2013; Kaya and Oren 2005) have been used as adsorbents for the removal of heavy metals from water and wastewater.

Among the natural clays of high surface area and eminent cation exchange capacity as well as the adsorption affinity for organic and inorganic ions and a low cost, bentonite is marked as the most promising adsorbent for the removal of heavy metals (arsenic, cadmium, chromium, iron, lead, zinc)

from aqueous solutions. Bentonite has a typical layered silicate structure consisting of two silica tetrahedral sheets with a central octahedral sheet. The interlayer space is easily accessible to water and another polar liquids. Bentonite has a good cation exchange capacity due to the presence of hydrated cations as Ca^{2+} , Na^+ , K^+ , etc., in their interlayer surfaces, and these cations can be easily exchanged by heavy metals compensating the negative charge.

There are many studies on the adsorption of iron and zinc (Sheta et al., 2003; Bhattacharyya and Gupta 2006; Mishra and Patel 2009) The aim of this work is to study the Macedonian natural bentonite as potential adsorbent for iron and zinc removal from wastewater.

EXPERIMENTAL

The bentonite samples in this study, originate from Kriva Palanka, (Ginevci district) Republic of Macedonia. Bentonite was used in its natural form without any treatment. For bentonite characterization, specific surface, sieve, chemical, XRD, FT- IR and TGA- DTA analysis were performed.

The initial metal ion concentration in the leachate taken from MSWL Drisla was determined by atomic absorption spectrophotometer (AAS) and it was $9,25 \text{ mg/dm}^3$ and $0,22 \text{ mg/dm}^3$ for Fe^{2+} and Zn^{2+} , respectively. The pH value of the leachate was 7,8. The batch method was used for adsorption processes. Volume of 25 cm^3 , 50 cm^3 , 75 cm^3 , 100 cm^3 and 125 cm^3 of leachate with 2 g of bentonite were placed in 5 Erlenmeyer flasks to commence the experiment. Flasks were set in the shaker (CERTOMAT R, B. Braun supplied by Biotech International) at constant temperature of 298 K. 100 rpm shaking was applied in the shaker for 5h and the last 1h at 140 rpm until equilibrium was attained. The solutions were filtered and the filtrates were analyzed for Fe^{2+} and Zn^{2+} by atomic absorption spectrophotometer.

RESULTS AND DISCUSSION

In this work specific surface of bentonite was evaluated from data of equilibrium adsorption of water vapor at 298K. The adsorption isotherm at 298K is interpreted by linear form of BET equation:

$$\frac{x}{n^a(1-x)} = \frac{1}{n_m^a C} + \frac{C-1}{n_m^a C} x \quad (1)$$

where, x is relative pressure ($x = P/P_0$), n^a [mol/g] is amount of adsorbed water vapour, n_m^a [mol/g] is monolayer capacity of adsorbent and C is a constant. The diagram of the linear form of the equation is given in Figure 1. The good fit of the experimental isotherm to the equation is evidenced by the linearity of the plot. The value of regression coefficient is higher than 0,98.

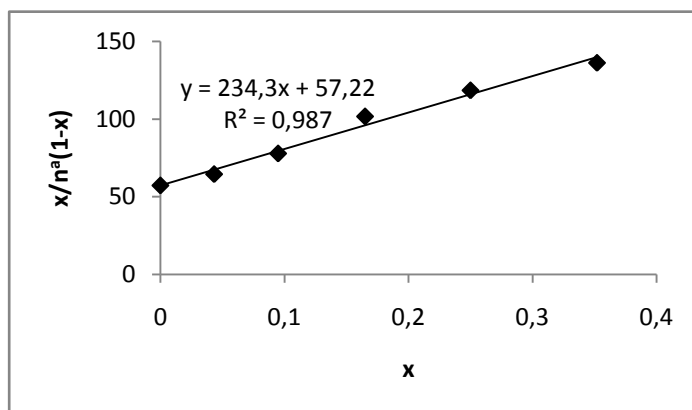


Figure 1. Diagram of linear form of BET equation of adsorption of water vapour on bentonite

The value of monolayer capacity is used to calculate the specific surface area of bentonite by the equation:

$$a_s = n_m^a N_A a_m 10^{-18} \quad (2)$$

where a_s is the specific surface area, a_m is surface of adsorbent occupied by one molecule of adsorbate, which for water is $0,106 \text{ nm}^2$.

The values of the parameters of the equation, monolayer capacity, n_m^a and constant C , and the specific surface area, a_s , are $3,43 \cdot 10^{-3} \text{ mol/g}$, $5,10$ and $219 \text{ m}^2/\text{g}$, respectively.

It was proved by XRD analysis on this bentonite, Figure 2, that the dominant component is montmorillonite and the remaining components are muscovite, quartz, feldspar and kaolinite.

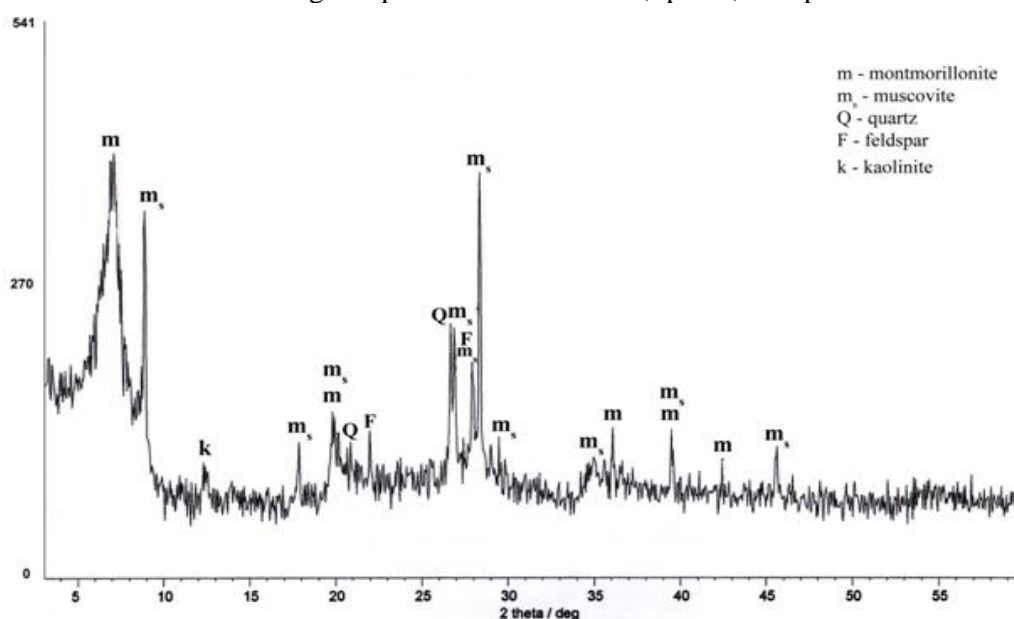


Figure 2. XRD patterns of the bentonite

Structural composition of B-KP revealed by XRD analysis was confirmed by FT-IR results, Figure 3. Namely, the adsorption bands at 467 , 517 , 783 and 1120 cm^{-1} point to the presence of quartz (Si-O-Si) while the bands at 1034 and 1634 cm^{-1} reveal the presence of SiO_2 amorphous. The bands appearing at 617 and 675 cm^{-1} outline the Al-O-Si-O bond (feldspar presence) and at 917 cm^{-1} the Al-Al-OH bond. The presence of adsorbed water is manifested by appearance of bands at 1634 , 3442 and 3617 cm^{-1} . OH groups considered with bands at 917 , 3442 and 3692 cm^{-1} were additionally confirmed by TGA-DTA analysis, Figure 4, that point to the different nature of OH groups present in B-KP.

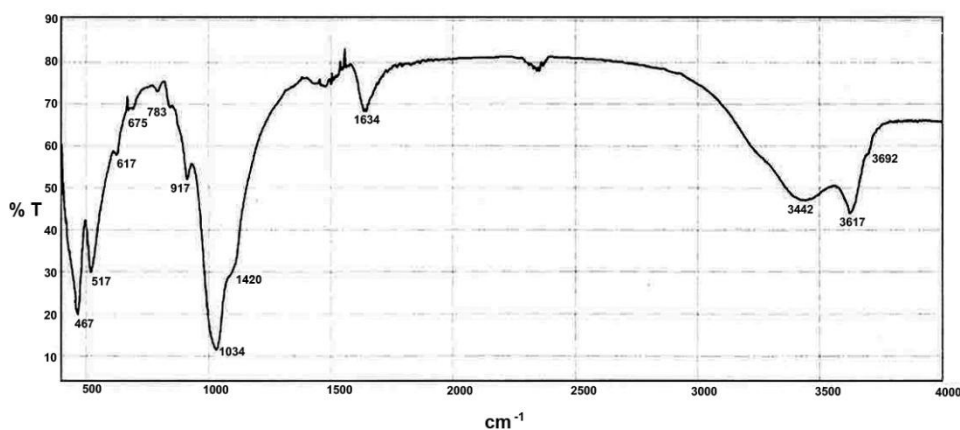


Figure 3. FT-IR spectrum of bentonite

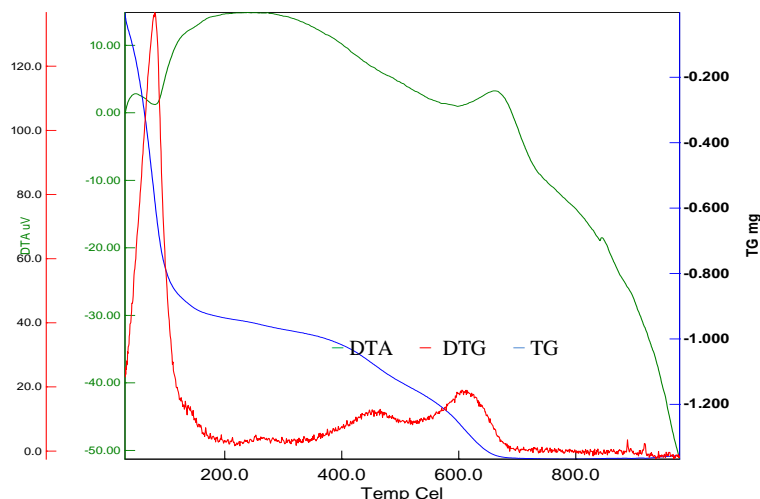


Figure 4. TGA-DTA diagram of bentonite

The results of granulometric sieve analysis and chemical composition of the bentonite determined by conventional silicate chemical analysis are presented in Table 1 and Table 2, respectively.

Table 1: Granulometric sieve analysis of bentonite

Fraction [mm]	Mass [%]
+0,071	3,82
-0,071+0,063	0,55
-0,063+0,040	5,26
-0,040+0,032	1,79
-0,032	88,58
Σ	100

Table 2: Chemical composition of bentonite

Composition	Mass [%]
SiO ₂	60,82
Al ₂ O ₃	21,32
Fe ₂ O ₃	3,21
MgO	4,15
CaO	2,03
Na ₂ O	0,40
K ₂ O	0,40
H ₂ O + CO ₂	8,06

Adsorption studies

The adsorbed amount of metal ion per unit adsorbent mass, m^a [mg/g], was calculated as follows:

$$m^a = \frac{(C_0 - C_e)V}{m} \quad (3)$$

where C_0 [mg/dm³] is the initial heavy metal concentration, C_e [mg/dm³] is equilibrium heavy metal concentration, m is the amount of the bentonite [g] and V is the leachate volume [dm³]. Initial concentrations of Fe²⁺ and Zn²⁺ metal ions in leachate were 9,25 mg/dm³ and 0,22 mg/dm³, respectively.

Adsorption isotherms are important for adsorption processes research. Langmuir and Freundlich isotherms are the most widely used for practical applications. The Langmuir adsorption isotherm is valid for monolayer sorption onto a surface with a finite number of identical sites (Melichova and Hromada 2013; Al Dwairi and Al Rawajfeh 2012), and it can be defined according to the following linear form:

$$\frac{C_e}{m^a} = \frac{1}{m_m^a b} + \frac{C_e}{m_m^a} \quad (4)$$

where m_m^a [mg/g] is the monolayer capacity and b is the equilibrium constant.

The dependence of C_e/m^a from C_e , obtained by using experimental results, is shown in Figure 5 for Fe^{2+} , and in Figure 6 for Zn^{2+} .

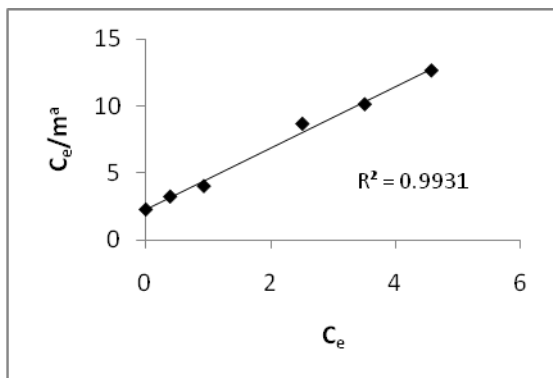


Figure 5. Langmuir adsorption model for Fe^{2+}

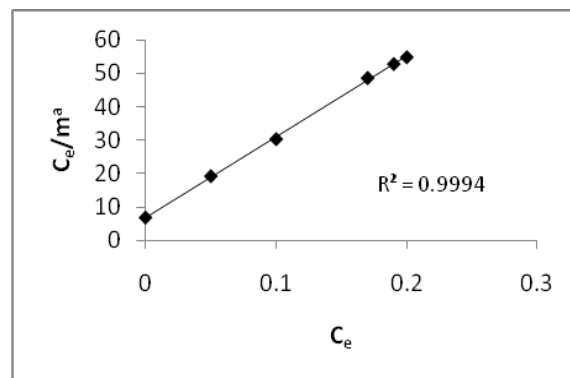


Figure 6. Langmuir adsorption model for Zn^{2+}

The values of m_m^a and b were determined from the linear form of the Langmuir equation and the Langmuir plot. The calculated results and correlation coefficients R^2 are listed in Table 3.

Langmuir isotherm determines the adsorption favorable or unfavorable. To determine the characteristic behavior of adsorption, dimensionless equilibrium parameter, R_L (known as the separation factor) is used (Al Dwairi and Al Rawajfeh 2012; Budsareechai et al., 2012). This parameter is given by:

$$R_L = \frac{1}{1 + bC_0} \quad (5)$$

For favorable adsorption R_L value must take place $0 < R_L < 1$. In our results R_L values were found 0,10 and 0,11 for Fe^{2+} and Zn^{2+} respectively, and that confirms that the adsorption process is favorable.

Freundlich isotherm is used for modeling the adsorption on heterogeneous surfaces. This isotherm can be explained by the linear form:

$$\log m^a = \log K_F + \frac{1}{n} \log C_e \quad (6)$$

where K_F is the Freundlich constant [mg/g] and n is an empirical parameter related to the intensity of adsorption. The linear Freundlich plots for iron and zinc are given in Figure 7 and Figure 8.

The parameters of Freundlich equation as well as the correlation coefficients for both metal ions are listed in Table 3.

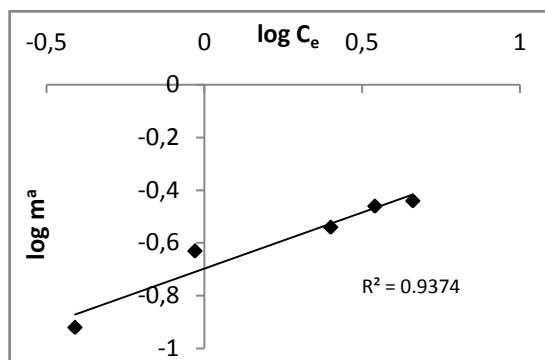
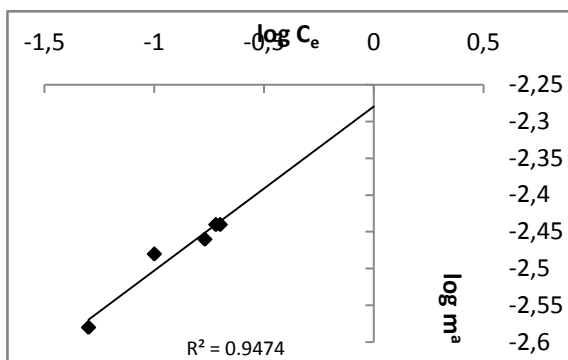

 Figure 7. Freundlich adsorption model for Fe^{2+}

 Figure 8. Freundlich adsorption model for Zn^{2+}

Table 3: Parameters of the Langmuir and Freundlich isotherms

Metal	Langmuir			Freundlich		
	m_m^a [mg/g]	b	R^2	K_F [mg/g]	N	R^2
Fe	0,4371	1,01	0,993	0,20	2,35	0,937
Zn	$4,1 \cdot 10^{-3}$	35,7	0,999	$5,25 \cdot 10^{-3}$	4,49	0,947

The linearity of the plots, Figures 5, 6, 7 and 8 and higher values of correlation coefficients show that Langmuir isotherms correspond better than Freundlich isotherms for both metals.

Langmuir monolayer capacity m_m^a and Freundlich constant K_F , that could be considered as the amount of metal adsorbed by solid phase at the equilibrium concentration of unity (Ghasemi-Fasae et al., 2012) do indicate and compare adsorption performance. According to the results for monolayer capacity, m_m^a and Freundlich constant K_F , bentonite shows greater sorption capacity for Fe^{2+} than for Zn^{2+} . According to many literature data (Budsareechai et al., 2012; Ghormi et al., 2013; Melichova and Hromada 2013) it was expected the bentonite to have higher values for monolayer capacity. In our case, because of the complexity of investigated system (leachate) which is highly polluted with many inorganic and organic pollutants and there is an increase of the competitiveness for bentonite adsorption sites, we have much lower results for sorption capacity.

The adsorption parameter n in the Freundlich isotherm which measures preferential adsorption of one adsorbate to other and the intensity of adsorption, compared directly with Langmuir constant b a factor which relate to heat of adsorption and affinity to the binding site (Okeola and Odeunmi 2010). As it can be seen from the Table 3 the two constants b and n are proportionally related for both metals and the higher values of b and n for Zn^{2+} indicate a stronger bond between bentonite and Zn^{2+} ions than bentonite and Fe^{2+} ions. Using mathematical calculations n values between 1 and 10 indicate effective adsorption (Al-Shahrani 2012) Our results show effective adsorption for Fe^{2+} and Zn^{2+} according to the n values which are 2,35 and 4,49 respectively.

From economic point of view it is necessary to define the optimal ratio between amount of used bentonite and dose of leachate. For that purpose the removal efficiency of Fe^{2+} and Zn^{2+} metal ions adsorption on bentonite was calculated by the following equation and the results are listed in Table 4.

$$\text{Removal \%} = \frac{C_0 - C_e}{C_0} \cdot 100 \quad (7)$$

Table 4: Removal efficiency of bentonite for Fe^{2+} and Zn^{2+}

Amounts of leachate/bentonite	Removal [%]	
	Fe	Zn
25 cm ³ / 2g	95,78	98,82
50 cm ³ / 2g	89,95	98,50
75 cm ³ / 2g	72,86	98,41
100 cm ³ / 2g	62,05	98,36
125 cm ³ / 2g	50,49	98,34

The results show that to keep removal percentage above 90% for Fe metal ions the ratio between amount of leachate and bentonite should be 25 cm³ / 2g because using the bigger volume of leachate drastically decreases the removal percentage. But, as it can be seen from the table, for Zn ions by increasing the leachate volume, percentage of removal, almost, do not change at all, and it is 98%.

CONCLUSION

The natural bentonite from Kriva Palanka, Republic of Macedonia (B-KP), belongs to the group of clay minerals with montmorillonite as main constituent. Bentonite has high specific surface area of 219 m²/g.

The linear form of Langmuir and Freundlich equations were used to describe the adsorption isotherms. The experimental results were well fitted with both isotherms, but the values of the correlation coefficients show that Langmuir model correspond better. The monolayer capacities of bentonite for iron and zinc were determined by Langmuir equation and are 0,4371 mg/g and 4,1·10⁻³ mg/g, respectively. This values are very low because we investigated a real system which is highly polluted and there is an increase of competitiveness for superficial sites in bentonite.

The results for removal percentage indicate that in that conditions bentonite is more effective as adsorbent for zinc and the removal efficiency above 98% is economically reasonable.

The further investigations would be conducted to examine adsorption characteristics of Macedonian bentonite under various operating variables like initial metal concentration, solution pH, contact time, clay dosage and temperature. Further studies would be dedicated to the investigations under which conditions the bentonite is no longer reusable as well as to determine the expected lifetime of the bentonite.

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